

**Appendix D****Dose Calculations**

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## D. Dose Calculations

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Dose calculations based on measured and/or estimated releases from stack emissions, liquid effluents, and contaminated soils were conducted for the public and biota. These dose calculations are summarized in Section 4.2. Details of the methods and assumptions used for modeling individual and population dose for the public are provided in Section D.1. Methods and assumptions related to the calculation of biota dose are provided in Section D.2.

### D.1 Supporting Information for Calculation of Public Doses

The radiological dose that the public could have received in 2015 from the Hanford Site was calculated in terms of the *total effective dose*. The total effective dose is the sum of the effective dose equivalent from external sources and the committed effective dose equivalent for internal exposure, which are summarized here and described in more detail in DOE's Occupational Radiation Protection rule in [10 CFR 835](#). The committed effective dose equivalent is the sum of doses to organs and tissues that is weighted to account for the sensitivity of the organ or tissue to the effects of radiation and for the biological effectiveness of the type of radiation causing the dose. It is expressed in units of rem (Sv), or more typically the sub-unit mrem (mSv)† for individuals, and in units of person-rem (person-Sv) for the collective dose received by the total population within a 50-mi (80-km) radius of Hanford Site operations areas. This appendix describes how the doses summarized in Section 4.2 of this report were calculated.

Calculation of the total effective dose accounts for the long-term (50 yrs) internal exposure from radionuclides absorbed into the body during the current year. The committed effective dose equivalent is the sum of individual committed (50 yrs) organ doses multiplied by tissue weighting factors (International Commission on Radiological Protection [ICRP] 1991) that represent the contribution of each organ or tissue to a person's internal radiation dose. Internal organs also may be irradiated from external sources of radiation. The external exposure received during the current year is added to the committed internal dose to obtain the total effective dose.

Releases of radionuclides from Hanford Site facilities are frequently too small for their concentrations to be accurately measured in many of the off-site environmental media of interest. Even when present in measureable amounts, it can be difficult to distinguish the small Hanford Site contributions from levels attributable to fallout from historical nuclear weapons testing and from naturally occurring radionuclides such as uranium and its decay products. Therefore, Hanford-related environmental radionuclide concentrations were estimated from stack effluent measurements (air pathway doses) or river water measurements (water pathway doses) by using environmental transport models. The air dose calculations employ environmental transport modeling based on measurements made at the points of release (stacks and vents). The water pathway dose calculations are based on the difference in measurements of radionuclide concentrations in the Columbia River upstream and downstream of the Hanford Site.

The transport of radionuclides in the environment to points of exposure is predicted using mathematical models of the physical processes underlying the various exposure pathways. These models are used to calculate radionuclide levels in air, soil, and foods at off-site locations. Long-lived radionuclides deposited

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†1 rem (0.01 Sv)=1,000 mrem (10 mSv).

on the ground by irrigation or airborne depositions become possible sources of external exposure and uptake by agricultural products. Radionuclides taken into the body by inhalation or ingestion may be distributed among different organs and tissues and retained in the body for various lengths of times. Agricultural, behavioral, and dosimetric models were applied to calculate radionuclide intakes and radiological doses to the public from annual-average radionuclide concentrations in the exposure media. Computer programs were used to implement these mathematical models using Hanford Site-specific dispersion and uptake parameters. These programs are incorporated in a master code—*GENII - The Hanford Environmental Radiation Dosimetry Software System, Version 2.10* ([PNNL-14583](#); [PNNL-14584](#); [PNNL-19168](#)), which employs the internal dosimetry methodology described in ICRP Publication 60 (ICRP 1991) and external dose coefficients described in Federal Guidance Report 12 ([EPA 1993](#)). GENII Version 1.485 ([PNL-6584](#), *The Hanford Environmental Radiation Dosimetry Software System*), which incorporated internal dosimetry methods of *International Commission on Radiological Protection Publication 30* (ICRP 1979a and 1979b) was used for dose calculations through 2008. GENII Version 2.10 is a Microsoft Windows®-based version that also incorporates some environmental modeling improvements (e.g., plume depletion during atmospheric transport) relative to Version 1.485. The modeling assumptions and radionuclide release data used in the GENII calculations are the primary focus of Section D.1. The ingestion and inhalation dose coefficients (ICRP 1991) and external dose coefficients ([EPA 1993](#)) used for the pathway dose calculations are described further in [PNNL-14584](#) and are not reproduced here.

In addition to the GENII calculations for assessing public doses, the computer program CAP88-PC (also known as CAP-88) was used to calculate an air pathway dose to a maximally exposed individual (MEI) for compliance with *Clean Air Act* standards, as required by the EPA through [40 CFR 61](#), Subpart H, from airborne radionuclide effluents (other than radon) released at DOE facilities. Air pathways calculations performed with the CAP-88PC computer code differ slightly from those performed in GENII. Technical details of the CAP88-PC calculations are provided in [DOE/RL-2016-10](#), *Radionuclide Air Emissions Report for the Hanford Site, Calendar Year 2015*.

Calculations of radiological doses to the public from radionuclides released into the environment are performed to demonstrate compliance with applicable standards and regulations. [DOE O 458.1](#) provides requirements for demonstrating compliance with the public dose limit of 100 mrem (1 mSv) total effective dose in a year. Relevant requirements include the following:

- Compliance may be demonstrated by calculating dose to the representative person or to the MEI
- Collective dose for members of the public should be calculated, and may be truncated, by distance (e.g., 50 mi [80 km])
- The representative person or MEI must include members of the public outside of controlled areas on DOE sites and off site
- Analytical models used to calculate dose must be codified or approved by DOE and must consider likely exposure pathways, including external radiation from air and soil, inhalation, and ingestion of water and terrestrial and/or aquatic foods
- Calculations of doses to the public from exposures resulting from both routine and unplanned activities must be performed using DOE-approved dose conversion factors
- Values of default or site-specific parameters used in the dose modeling must be included to document the calculations.

A summary of how the location of the off-site MEI was identified, and information on modeling assumptions and inputs to the GENII computer code used to conduct the MEI dose calculations is provided in Section D.1.1. Information supporting the calculation of collective off-site dose for members of the public using the GENII computer code is provided in Section D.1.2.

### D.1.1 Maximally Exposed Individual Dose

The MEI is a hypothetical member of the public whose location and lifestyle make it unlikely that any actual individuals would receive higher doses. The location of the MEI can vary annually depending on 1) the relative contributions of the different operational areas to radioactive emissions released to the air, 2) the contribution of radionuclide releases to the Columbia River from Hanford Site facilities, and 3) variable differences in meteorology affecting wind dispersion. The following potentially significant exposure pathways are considered for identifying the location of this hypothetical individual and calculating radiation dose:

- Inhalation of airborne radionuclides
- External exposure from submersion in airborne radionuclides
- Ingestion of foodstuffs contaminated by radionuclides deposited on vegetation and the ground by airborne deposition and/or irrigation water drawn from the Columbia River downstream of the Hanford Site
- Incidental ingestion of soil and external exposure to ground contaminated by airborne deposition and/or irrigation water
- Ingestion of drinking water drawn from the Columbia River
- Consumption of fish from the Hanford Reach of the Columbia River
- Recreational activities along the Hanford Reach of the Columbia River, including fishing, hunting, boating, swimming, and exposure to sediments during shoreline activities.

**Determination of the Location of the MEI.** Based on experience since 1990 from environmental transport modeling and environmental surveillance monitoring, four locations (Section 4, Figure 4.2) are considered for identifying the location of the MEI. The distinguishing characteristics of these locations are described in the following paragraphs.

**Riverview MEI.** The Riverview area is across the Columbia River from the City of Richland. Because of its location, an individual in the Riverview area has the potential to receive the maximum exposure to waterborne effluent from Hanford Site facilities as well as some contribution from exposure to airborne emissions from the 300 Area. The Riverview location is where a small population of West Pasco residents obtain their drinking water from the river via a community water system; therefore, the domestic drinking water pathway is applied to this location. Columbia River water from just downstream of the Hanford Site is also withdrawn for irrigation of small gardens and farms at Riverview.

**Ringold MEI.** The Ringold area is along the eastern shoreline of the Columbia River 16 mi (26 km) east of separations facilities in the 200 Areas. Because of its location, an individual in the Ringold area has the potential to receive the maximum exposure to airborne emissions from the 200 Areas. In addition, it is assumed that some individuals in the Ringold area may irrigate their crops with water from the Columbia River downstream of where contaminated groundwater originating from the 100 and 200-East Areas enters the river. For identifying the MEI, Hanford Site contributions to irrigation water at Ringold are

protectively evaluated using the same downstream concentrations employed for Riverview. Domestic drinking water at Ringold is not obtained from the Columbia River so this exposure pathway is incomplete.

*Sagemoor MEI.* An individual in the Sagemoor area, located 0.87 mi (1.4 km) directly across the Columbia River from the 300 Area, frequently receives maximum exposure to airborne emissions from the 300 Area. However, domestic water at this location comes from wells rather than from the river; as a result, wells on the eastern side of the Columbia River are not impacted by radionuclides of Hanford Site origin. Because the farms located across from the 300 Area obtain irrigation water from the Columbia River upstream of the Hanford Site, irrigation-related exposure pathways are likely incomplete at this location. However, because some individuals may obtain much of their food from local agriculture, Columbia River irrigation pathways agricultural dose has been historically assigned to the Sagemoor area MEI. This practice protectively but unrealistically sums the location-specific air deposition component of food-related dose with the irrigation component from another location. The added contribution of radionuclides in the Riverview area irrigation water maximizes the calculated dose from the air and water pathways combined.

*Horn Rapids Road MEI.* Meteorological conditions in 2012 through 2015 resulted in a more southerly direction of wind dispersion than has been observed in past years. As a result, air concentrations related to 300 Area emissions were modeled to be slightly higher at a location just to the south of the Hanford Site boundary than at the Sagemoor location across the Columbia River to the east. Buildings in this area historically have been associated with commercial and industrial activities. However, in recent years, residences also have been constructed near the southern boundary of the Hanford Site south of the 300 Area. Residences in this area obtain drinking water from the City of Richland, which has an intake on the Columbia River downstream of the Hanford Site; therefore, the domestic drinking water pathway is applied to this location. Additionally, some agriculture in this area occurs on leased property that receives irrigation water from the Battelle pumping station on the Columbia River just below the 300 Area.

During the period of plutonium production at the Hanford Site, Ringold was commonly the location of the MEI. Because of the shift in Hanford Site operations from nuclear weapons production to the current mission of managing waste products, cleaning up legacy waste, and researching new ideas and technologies for waste disposal and cleanup, the significance of air emissions from production facilities in the 200 Areas has decreased compared to emissions from research facilities in the 300 Area. For the past two decades, the hypothetical MEI has been associated with air emissions from the 300 Area.

Because the hypothetical MEI at all locations is assumed to potentially receive dose from consumption of foods raised using Columbia River irrigation water, the identification of the location of the MEI is based on the highest projected dose among the following air pathways: at Ringold (200 Area sources), Sagemoor (300 Area sources), Horn Rapids Road (300 Area sources) plus drinking water pathways dose, and Riverview (300 Area sources) plus drinking water pathways dose.

For 2015, air-pathways radiological dose calculations conducted using CAP88-PC in support of *Clean Air Act* requirements and GENII Version 2.10 have identified Horn Rapids Road as the location with the highest MEI dose. Air pathways calculations performed with the GENII computer code indicate that Sagemoor and Horn Rapids Road air pathways MEI doses in 2015 are similar (0.12 mrem at Sagemoor and 0.15 mrem at Horn Rapids Road). Unlike the Sagemoor receptor the MEI at Horn Rapids Road receives additional dose from the drinking water pathway. Both Sagemoor and Horn Rapids Road MEI GENII results are shown in Section 4, Figure 4.4 for comparison.

MEI location coordinates relative to Hanford Site operating areas are entered in the GENII computer code to specify the location for the air pathways dose calculations. For Sagemoor, these coordinates are:

100 Area: 26.874 km Easting, 30.064 km Northing	300 Area: 1.35 km Easting, 0.26 km Northing
200 Area: 24.954 km Easting, 20.814 km Northing	400 Area: 7.909 km Easting, 6.739 km Northing

For Horn Rapids Road, these coordinates are:

100 Area: 29.1 km Easting, -29.1 km Northing	300 Area: 0 km Easting, -1.80 km Northing
200 Area: 22.6 km Easting, -22.6 km Northing	400 Area: 7.92 km Easting, -7.92 km Northing

**Water and Air Release Inputs Used In GENII Version 2.10.** As discussed in Section 4.2, the environmental data needed to perform the GENII dose calculations for the water pathways are differences in the measured upstream and downstream radionuclide concentrations in the Columbia River. The radionuclide releases to the Columbia River that are assumed to be the source of these differences are assigned to the 200 Areas, but area assignment does not affect the dose results. Measured emissions of radionuclides in stack releases are used in the GENII air pathways dose calculations. These air and water pathways data must be processed for input to the GENII computer code. GENII accepts inputs for environmental releases using dimensions of activity (e.g., Cu or Bq) per time for both water and air pathways.

Direct liquid effluent releases from outfalls in the 100 Areas were historically used to characterize contributions from the 100 Areas. The last operating outfall, 1908-K in the 100-K Area, ceased operations at the end of March 2011; therefore, no annual releases were identified from the 100 Areas in 2015. Liquid effluent discharges related to historical Hanford operations are known to enter the Columbia River by groundwater discharge at certain locations along the site shoreline from the 100-B/C Area downstream to the 300 Area. The impact of these discharges was evaluated as the difference between near-shore river water radionuclide concentrations downstream of the Hanford Site (monthly samples collected at the Richland Pumphouse, sampling location label RICH.PMPHS HRM46.4) and upstream samples collected below the Priest Rapids Dam (monthly samples collected at sampling location label PRIEST RAPIDS-RIVER). Some radionuclides are measured in both filtered samples (in solution) and in samples that capture suspended particulates (adhered to resin). These data were evaluated both separately and summed.

One-tailed paired t-tests and nonparametric Wilcoxon Rank Sum (WRS) tests were used to determine whether average downstream sample concentrations were statistically greater than upstream average concentrations. The results of statistical tests were used in conjunction with supporting information such as known releases from groundwater plumes to the river and historical observations in river water to identify Hanford-related contaminants. The paired t-test is more powerful than the ordinary t-test when the values in the pairs correlate, or when the concentrations measured downstream tend to correlate to those upstream. The WRS test has less power than the t-test when the data originate from a normal distribution, but the assumptions under which the statistical results are valid are not as restrictive. A p-value of 0.05 is commonly used as the threshold of statistical significance, but a larger, less restrictive value may be used when other factors support evidence of a release. Likewise, when a contaminant does not have any known Hanford Site sources a smaller threshold could be appropriate to identify radionuclides for the dose assessment.

Both statistical tests identified tritium and uranium-238 as potentially Hanford-related contaminants to include in the 2015 water-pathways dose assessment using a p-value of 0.05. Concentrations of uranium-

234 were greater downstream, but p-values were slightly higher than 0.05 for both the t-test and WRS. Uranium-234 is retained as a potentially Hanford-related contaminant for the 2015 dose assessment because the higher downstream concentrations are considered plausibly site-related, particularly in light of the uranium-238 results. Although uranium-235 might be expected to co-occur with both uranium-234 and uranium-238, yearly average uranium-235 concentrations were higher upstream than downstream and therefore uranium-235 is not included in the water pathways dose assessment calculations. Two additional radionuclides included in the 2015 water pathways dose calculations are cesium-134 and strontium-90. Neither radionuclide was measured at concentrations above detection limits. In both cases, the t-test p-values were below 0.05 and the WRS p-values were slightly greater than 0.05. Strontium-90 is known to be a component of groundwater plumes impacting the Columbia River and is included for that reason, even though it was not detected in the river water samples. Current Hanford Site-related cesium-134 releases have not been identified to the Columbia River, and its relatively short half-life of approximately 2 years suggests that its presence related to historical Hanford operations is unlikely. In addition, cesium-134 was not detected in any samples; nevertheless, it has been protectively included in the water pathways dose calculations.

	paired t-test	p-value Wilcoxon Rank Sum
Tritium	0.0000060	<0.0003
Uranium-238	0.0020	0.0067
Uranium-234	0.070	0.057
Cesium-134	0.023	0.057
Strontium-90	0.033	0.065

Table D.1 summarizes the mean annual differences in downstream and upstream concentrations, and calculated annual releases for the 2015 GENII water pathways dose calculations.

**Table D.1. Liquid Effluent Radionuclide Releases for GENII Dose Calculations**

Radionuclide	Upstream	Downstream	Difference
<b>Columbia River Annual-Average Radionuclide Concentrations (pCi/L)*</b>			
Cesium-134	1.2E-03	3.1E-03	1.9E-03
Strontium-90	1.1E-04	1.7E-02	1.7E-02
Tritium	1.7E+01	3.0E+01	1.3E+01
Uranium-234	2.5E-01	2.7E-01	2.0E-02
Uranium-238	1.9E-01	2.2E-01	2.6E-02
<b>Calculated Radionuclide Releases (Ci/year)†</b>			
Cesium-134	NA	NA	1.9E-01
Strontium-90	NA	NA	1.7E+00
Tritium	NA	NA	1.3E+03
Uranium-234	NA	NA	2.1E+00
Uranium-238	NA	NA	2.6E+00
Thorium-234‡	NA	NA	2.6E+00
Protactinium-234m‡	NA	NA	2.6E+00
NA=not applicable; radionuclide releases calculated based on difference between annual-average downstream and upstream concentrations.			
*1 pCi=0.037 Bq			
†Calculated as the product of the difference in downstream and upstream radionuclide concentrations and the 2015 annual-average river flow rate of 3,177 m <sup>3</sup> /sec at Priest Rapids Dam and the number of seconds in a year.			
‡These short-lived progeny of uranium-238 were protectively assumed to be in secular equilibrium at the time of discharge. Refer to Section 7.0 for information on Columbia River surface water sampling.			

Radioactive air emissions based on monitoring of stacks in the 100, 200, 300, and 400 areas were used as the basis for the GENII air pathways dose calculations. Stack emissions are measured for specific radionuclides related to the operations at each emissions point. During the dispersion time from the stack to an off-site exposure location there is opportunity for ingrowth of short-lived radioactive progeny that are included in the GENII radionuclide inventory. A protective upper-bound dispersion time of 15 hrs was estimated based on the longest dispersion distance in the collective dose calculations 50 mi (80 km) and an assumed (4.9 ft/sec [1.5 m/sec]) average wind speed. The highest short-term (15-hr ingrowth period) concentrations of short-lived progeny that have a separate dose conversion factor were included in the GENII air emissions inventory to address their potential contribution to the inhalation dose. Ingrowth of longer-lived progeny in soil and other environmental media is accounted for within GENII.

In addition to measurement of specific radionuclides, gross alpha and gross beta measurements are also made on emissions from each operating area. Following the precedent of the *Radionuclide Air Emissions Report for the Hanford Site, Calendar Year 2015* (DOE/RL-2016-10), measurements of gross alpha and gross beta radiation in stack emissions were protectively added to the measured emissions of plutonium-239/-240 and cesium-137, respectively, to ensure that contributions from any unmeasured operations-related radionuclides are incorporated in the estimated doses. These specific radionuclides were selected based on their historical association with releases in these operating areas and because air pathways calculations indicate dose is highest for these radionuclides among the group of plausible candidates of alpha- and beta-emitting radionuclides. Annual radionuclide air releases used in the GENII air pathways dose calculations are summarized in Table D.2.

**Table D.2. Air Pathways Radionuclide Stack Emissions for GENII Modeling**

Radionuclide	100 Area	200 Areas	300 Area (Ci)	400 Area
Hydrogen-3 (elemental tritium)	NA	NA	133	NA
Hydrogen-3 (tritiated water vapor)	NA	NA	282	1.8E-03
Carbon-14	NA	NA	1.2E-04	NA
Sodium-22	NA	NA	NA	1.4E-09
Krypton-85	NA	NA	5.8E-07	NA
Stontium-90	2.3E-06	1.0E-04	2.2E-07	NA
Yttrium-90 <sup>†</sup>	3.4E-07	1.5E-05	3.3E-08	NA
Technetium-99	NA	NA	4.1E-06	NA
Ruthenium-106	NA	2.6E-06	1.9E-09	NA
Iodine-129	NA	2.3E-04	NA	NA
Cesium-134	NA	NA	NA	NA
Cesium-137 <sup>‡</sup>	1.7E-05	2.9E-04	7.2E-06	1.3E-06
Barium-137m <sup>†,‡</sup>	1.7E-05	2.9E-04	7.2E-06	1.3E-06
Europium-152	NA	NA	1.8E-09	NA
Europium-154	NA	NA	1.2E-08	NA
Gadolinium-153	NA	NA	7.0E-11	NA
Radon-219	NA	NA	6.2E+00	
Lead-211 <sup>†</sup>	--	--	1.1E-02	
Bismuth-211 <sup>†</sup>	--	--	1.9E-03	
Thallium-207 <sup>†</sup>	--	--	8.7E-05	
Radon-220	NA	NA	385	NA
Lead-212 <sup>†</sup>	--	--	5.5E-01	NA
Bismuth-212 <sup>†</sup>	--	--	4.6E-01	NA
Radon-222	NA	NA	NA	NA

**Table D.2. Air Pathways Radionuclide Stack Emissions for GENII Modeling**

Radionuclide	100 Area	200 Areas	300 Area	400 Area
Radium-226	NA	NA	4.1E-10	NA
Actinium-227	NA	NA	1.8E-09	NA
Uranium-232	NA	NA	5.4E-09	NA
Uranium-233	NA	NA	1.8E-08	NA
Neptunium-237	NA	NA	1.4E-08	NA
Plutonium-238	5.1E-07	5.4E-07	3.8E-08	NA
Plutonium-239/240§	1.5E-05	6.3E-05	7.8E-08	2.5E-07
Plutonium-241	2.0E-05	1.6E-05	6.7E-07	NA
Americium-241	3.4E-06	4.6E-06	1.2E-09	NA
Americium-243	NA	NA	4.5E-08	NA
<i>Neptunium-239†</i>	--	--	7.6E-09	NA
(gross alpha)	1.1E-05	4.0E-05	7.2E-08	2.5E-07
(gross beta)	1.3E-05	2.5E-04	5.3E-06	1.3E-06

NA=Not available or not detected. No stack emissions reported for this radionuclide.

\*Radionuclides in *italic font* are short-lived progeny of the parent listed above that may ingrow during air dispersion to off-site locations.

†Values of these short-lived progeny are the highest activity calculated within an estimated 15-hr dispersion time period to an exposure point within a 50-mi (80-km) distance.

‡Values include the addition of gross beta activity.

§Values include the addition of gross alpha activity.

**Exposure Parameter Values Used in GENII Version 2.10.** GENII Version 2.10 requires input values for numerous parameters used in the environmental transport and human exposure models. Important parameters affecting the movement of radionuclides within agricultural exposure pathways such as animal dietary parameters, irrigation rates, crop yield, growing periods, and holdup periods are listed in Table D.3. The plant, animal, and aquatic foods transfer factors used for the pathway dose calculations are documented in [PNNL-14584](#) and are not reproduced here.

The off-site radiological dose is related to the extent of external exposure to or intake of radionuclides released from Hanford Site operations that become incorporated in exposure media such as air, water, soil, sediment, and various foodstuffs. Tables D.4 through D.6 provide the values for the diet, residency, and river recreation parameters for the MEI and collective dose (average individual) calculations.

**Meteorological Data Used in GENII Version 2.10.** GENII Version 2.10 employs an atmospheric dispersion model to calculate annual-average air concentrations and deposition rates at downwind locations based on site-specific radionuclide air emissions measurements and meteorological data ([PNNL-14583](#)). The 2015 meteorological data used in the GENII air dispersion modeling were gathered at monitoring stations in the 100 Area (station 13; 100-N), 200 Area (station 21; Hanford Meteorological Station), 300 Area (Station 11; 300 Area), and 400 Area (station 9; Fast Flux Test Facility [FFTF]). With the exception of the 200 Area, all meteorological data were obtained at a height of 33 feet (10 m). In the 200 Area, where some active stacks are 200 ft (61 m) in height, the meteorological data used were collected at 200 ft (61 m).

**Table D.3.** Agricultural Pathway Parameters for Hanford Site Dose Calculations

Medium	Vegetables							Milk	Hay (beef cattle, milk cows)	Pasture (milk cows)	Grains (beef cattle, poultry)
	Leafy	Root	Fruits	Cereals	Eggs	Poultry	Beef				
Holdup time; day (MEI)	1	5	5	180	1	1	15	1	100	0	180
Holdup time; day (population)	14	14	14	180	18	34	34	4	100	0	180
Growing period; day	90	90	90	90	NA	NA	NA	NA	45	30	90
Yield; kg/m*	1.5	4	2	0.8	NA	NA	NA	NA	2	1.5	0.8
Irrigation rate; cm/yr	77	88	77	†	NA	NA	NA	NA	103	103	†
Irrigation period; month	6	6	6	†	NA	NA	NA	NA	6	6	†
Water intake; L/year	NA	NA	NA	NA	0.3	0.3	50	60	NA	NA	NA
Food intake; kg/day	NA	NA	NA	NA	0.12	0.12	68/68‡	55/55§	NA	NA	NA
Contaminated fraction of diet*	NA	NA	NA	NA	1.0	1.0	0.25/0.75‡	0.25/0.75§	NA	NA	NA
Livestock soil intake; kg/day	NA	NA	NA	NA	0.0	0.0	0.0	0.375**	NA	NA	NA

Holdup=time between harvest and consumption; MEI=maximally exposed individual; NA=not applicable  
 \*Pertains to animal feed; 100% of animal water is assumed contaminated surface water.  
 †No irrigation is assumed to occur for cereal crops or grains.  
 ‡First value pertains to grains, and second value pertains to hay.  
 §First value pertains to hay, and second value pertains to pasture grass.  
 \*\*Calculated as 0.5 kg soil/day while grazing × 0.75 diet fraction of pasture grass.

**Table D.4. Consumption Parameters for Hanford Site Dose Calculations**

Medium	Consumption Rate*	
	Maximally Exposed Individual	Average Individual (Collective Dose)
Leafy vegetables	66 lbs (30 kg)/yr	33 lbs (15 kg)/yr
Root vegetables	485 lbs (220 kg)/yr	310 lbs (140 kg)/yr
Fruits	728 lbs (330 kg)/yr	140 lbs (64 kg)/yr
Cereals	180 lbs (80 kg)/yr	160 lbs (72 kg)/yr
Milk	71 gal (270 L)/yr	61 gal (230 L)/yr
Beef	180 lbs (80 kg)/yr	150 lbs (70 kg)/yr
Poultry	40 lbs (18 kg)/yr	19 lbs (8.5 kg)/yr
Eggs	66 lbs (30 kg)/yr	44 lbs (20 kg)/yr
Fish†	88 lbs (40 kg)/yr	--‡
Drinking water§	193 gal (730 L)/yr	116 gal (440 L)/yr
Inadvertent soil ingestion	1.17 oz (36.5 g)/yr	0.59 oz (18.3 g)/yr

\*A transit time of 11 hrs from the release to receptor locations is assumed.

†A holdup time of 1 day is used for both MEI and population calculations.

‡Average individual consumption not identified; see text of Section D.1.2.

§A holdup time of 1 day is used for the Riverview calculations for identification of the location of the MEI.

**Table D.5. Residency Parameters for Hanford Site Dose Calculations**

Pathway	Exposure	
	Maximally Exposed Individual	Average Individual (Collective Dose)
Air: Inhalation*, †	24 hrs/day, 365 days/yr	24 hrs/day, 365 days/yr
Air: external (submersion)†	24 hrs/day, 365 days/yr	24 hrs/day, 365 days/yr
Soil: external (ground shine)	12 hrs/day, 365 days/yr	8 hrs/day, 365 days/yr

\*Inhalation rate, adult 1.0 m<sup>3</sup>/hr (35 ft<sup>3</sup>/hr).

†Dispersion time of 15 hrs is protectively assumed for ingrowth of short-lived progeny during transport (50 mi [80 km]) population dose radius and 4.9 ft/s (1.5 m/s) wind speed.

**Table D.6. Columbia River Parameters for Hanford Site Dose Calculations**

Activity and Pathway	Exposure*	
	Maximally Exposed Individual	Average Individual (Collective Dose)
Shoreline: sediment; external	5.0 hrs/day, 100 days/yr†	1.7 hrs/day, 10 days/yr†
Boating: river water; external	2.0 hrs/day, 50 days/yr‡	0.1 hr/day, 50 days/yr‡
Swimming: river water; inadvertent ingestion§, external	2.0 hrs/day, 50 days/yr	0.2 hr/day, 50 days/yr

\*A transit time of 11 hrs from the release to receptor locations is assumed.

†A shoreline width factor of 0.2 is used.

‡No shielding by the boat is assumed.

§Ingestion rate of 0.68 oz (0.02 L)/hr.

Because meteorological station 29 (100-K) has been inoperable since 2013, a meteorological file was compiled with data for 2003–2012, which represents a 10-yr average of daily meteorological data at this location. The 100 Area air pathways doses were calculated using this 10-yr average data file and using 2015 meteorological data from Station 13 at 100-N. The difference in dose results was approximately 7%. The 2015 100-N meteorological file, which is more applicable to local and regional air dispersion in 2015, was used for the MEI and population dose calculations. The negligible difference in dose using the 100-K

10-year average meteorological data file versus the 2015 100-N air station file indicates the selection of one or the other has little impact on the dose assessment results.

Hourly meteorological data from the monitoring stations described above were formatted for use in the GENII computer code. Five meteorological files, one for each of the Hanford Site operating areas and stations described above, were created. These files were referenced in the GENII Chronic Plume Air Module. A radial grid consisting of 16 directional sectors and 10 downwind distances was specified in the air module. The downwind distances were varied for each operating area to coincide with the distance to the MEI location, as defined by the Easting and Northing coordinates described above. For example, the finest resolution was entered for the distance from the 300 Area to the MEI location

### **D.1.2 Fifty-Mile (Eighty-Kilometer) Collective Population Dose**

Regulatory limits have not been established for collective doses to a population; however, evaluation of the collective population doses to all residents within a 50-mi (80-km) radius of Hanford Site operations is required by [DOE O 458.1](#). The radiological dose received by the total population within 50 mi (80 km) of site operations areas was calculated to confirm adherence to DOE environmental protection policies, and provide information to the public. The 50-mi (80-km) collective dose is the sum of doses to all individual members of the public within 50 mi (80 km) of the four Hanford Site operations areas (100 Area, 200 Area, 300 Area, and 400 Area).

The same exposure pathways evaluated for the MEI (Section D.1.1) were used to calculate doses to the off-site population. The primary difference between the MEI and collective dose calculations is in the values selected for certain exposure parameters. As shown in Tables D.4, D.5, and D.6, exposure parameter values for the collective dose calculations reflect an average individual rather than an MEI.

In calculating the collective dose related to water-mediated exposure pathways (drinking water, irrigated foods, Columbia River recreation, and fish consumption), estimates were made of the off-site population size expected to be affected by each pathway. The assumptions of population size and the calculation of collective dose for each of these four exposure pathways are described in the following paragraphs.

**Drinking Water.** The Cities of Richland and Pasco obtain all or part of their municipal water directly from the Columbia River downstream from the Hanford Site; the City of Kennewick obtains its municipal water indirectly from wells adjacent to the river. Approximately 182,000 people residing in the Tri-Cities‡ are assumed to obtain all of their drinking water directly from the Columbia River or from impacted wells near to the river. Annual drinking water dose for an average individual is multiplied by the Tri-Cities population to calculate the collective drinking water dose.

**Irrigated Food.** Columbia River water is withdrawn for irrigation of small vegetable gardens and farms in the Riverview area of Pasco in Franklin County. It is assumed enough food is grown in this area to feed an estimated 2,000 people. Commercial crops are also irrigated by Columbia River water in the Horn Rapids area of Benton County. Because these crops are widely distributed, any individual in the Tri-Cities is likely to receive only negligible potential exposure. Annual irrigated foods dose for an average individual is multiplied by the estimate population of 2,000 individuals to calculate the collective irrigated foods dose.

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‡The Cities of Pasco, Kennewick, and Richland—known as the Tri-Cities—are located in southeastern Washington State. Population estimates from <http://quickfacts.census.gov/qfd/index.html>.

**Columbia River Recreation.** As described in Section 4.2 and Section D.1.1, these recreational activities include fishing, swimming, boating, and shoreline recreation. It was protectively assumed that all 182,000 individuals in the Tri-Cities participated in these recreational activities. Annual recreational dose for an average individual is multiplied by the Tri-Cities population to calculate the collective recreational dose.

**Fish Consumption.** Population doses from consuming fish obtained locally from the Columbia River were calculated from an estimated total annual catch of 33,000 lbs (15,000 kg) per year. It was protectively assumed that 100 percent of the annual catch was consumed by individuals in the Tri-Cities area. Population dose related to fish consumption was calculated as follows:

$$\text{Population dose (person-rem)} = \text{MEI dose (mrem)} \times 0.001 \text{ rem/mrem} \times (\text{annual catch [kg/yr]}/\text{IR\_fish [kg/yr/person]}), \text{ where MEI dose=fish ingestion dose for the MEI; annual catch=15,000 kg fish/yr; IR\_fish=individual fish ingestion rate used in the MEI calculation (40 kg/yr/person).}$$

Collective dose related to air-mediated exposure pathways was calculated based on the geographic distribution of the population residing within a 50-mi (80-km) radius of the Hanford Site operating areas, as shown in [PNNL-20631](#). These distributions are based on 2010 Bureau of the Census data and influence the population dose by providing estimates of the number of people exposed to radioactive air emissions and their proximity to the points of release.

The air pathways collective dose calculations are based on modeled radionuclide air concentrations and deposition rates downwind of the Hanford Site operating areas coupled with the geographic population distribution in these areas. Both meteorological and population distribution data are organized according to 16 directional sectors based on the four cardinal, four ordinal, and eight cross-wind directions (N, NNE, NE, ENE, etc.). These sectors were transformed into grids using concentric circles with radii of 1, 2, 3, 4, 5, 10, 20, 30, 40, and 50 mi. These radii correspond to the downwind distances specified in the GENII Chronic Plume Air Module. Population files were created based on the number of individuals located in each of the 160 grid segments centered on the 100, 200, 300, and 400 areas ([PNNL-20631](#)). These files were identified in the GENII Air Dose Report Module.

## D.2 Calculation of Biota Doses

The RESRAD-BIOTA computer code was used to screen the 2015 radionuclide concentrations in water, sediment, soil, and tissues to see if they exceeded the established biota concentration guides. Biota concentration guides are concentrations published in *A Graded Approach for Evaluating Radiation Doses to Aquatic and Terrestrial Biota* that could result in a dose rate of 1 rad per day for aquatic biota or 0.1 rad per day for terrestrial organisms. Table D.7 presents water and sediment Tier 1 biota concentration guides for the radionuclides evaluated. Table D.8 presents the soil Tier 1 biota concentration guides for the radionuclides evaluated. Both internal and external doses to aquatic, riparian, and terrestrial animals and plants are included in the screening process. For with multiple media and radionuclides analyses, a sum of fractions is calculated to account for the contribution to dose from each radionuclide relative to its corresponding biota concentration guide. If the sum of fractions exceeds 1.0, then the dose guideline has been exceeded. If the initial estimated screening value (Tier 1) exceeds the dose limit (sum of fractions more than 1.0), additional screening calculations are performed (Tiers 2 or 3) to evaluate accurate exposure of biota to radionuclides. The process may culminate in a site-specific assessment requiring additional sampling and study of exposure.

**Table D.7. Biota Concentration Guides and Sediment to Water Partition Coefficients**

Radionuclide	Water (pCi/L)*	Limiting Organism	Sediment (pCi/g)*	Limiting Organism	Default Kd (mL/g)†
Hydrogen-3	2.65E+08	Riparian animal	3.74E+05	Riparian animal	0.001
Carbon-14	6.09E+02	Riparian animal	5.90E+04	Riparian animal	0.001
Cobalt-60	3.76E+03	Aquatic animal	1.46E+03	Riparian animal	1000
Strontium-90	2.78E+02	Riparian animal	5.82E+02	Riparian animal	30
Technetium-99	6.67E+05	Riparian animal	4.22E+04	Riparian animal	5
Cesium-137	4.26E+01	Riparian animal	3.12E+03	Riparian animal	500
Plutonium-238	1.76E+02	Aquatic animal	5.73E+03	Riparian animal	2000
Plutonium-239/-240	1.87E+02	Aquatic animal	5.86E+03	Riparian animal	2000
Uranium-234	2.02E+02	Aquatic animal	5.27E+03	Riparian animal	50
Uranium-235	2.17E+02	Aquatic animal	3.73E+03	Riparian animal	50
Uranium-238	2.23E+02	Aquatic animal	2.49E+03	Riparian animal	50

\*1 pCi=0.037 Bq. Biota concentration guides (pCi/g or pCi/L) from RESRAD-BIOTA v1.5.

†Kd=Water to Sediment Partition Coefficients (mL/g) from RESRAD-BIOTA v1.5.

**Table D.8. Tier 1 Soil Biota Concentration Guides**

Radionuclide	Soil (pCi/g)*	Limiting Organism
Strontium-90	2.25E+01	Terrestrial animal
Cesium-137	2.08E+01	Terrestrial animal
Plutonium-238	5.27E+03	Terrestrial animal
Plutonium-239/240	6.11E+03	Terrestrial animal

\*1 pCi=0.037 Bq. Biota concentration guides (pCi/g) from RESRAD-BIOTA v1.5.

In the initial (Tier 1) screening assessment, researchers compare maximum measured concentrations to the biota concentration guides. The maximum detected concentrations evaluated for aquatic biota dose assessment are presented in Table D.9. If the sum of fractions does not exceed one, no further analysis is required. However, if the sum of fractions does exceed one, a second analysis (Tier 2) is performed using average concentrations and the same Tier 1 biota concentration guides. For the aquatic biota dose assessment, paired sediment and water data are required. In the event that only one of these media was sampled, the other was calculated using an element-specific water to sediment partition coefficient. These coefficients are tabulated in Table D.7.

The sites for the aquatic biota dose assessment were grouped into upstream (Vernita sediment and river water samples), the 100 Areas (sediments from 100-K, 100-D, 100-F and White Bluff; river water from 100-K and 100-N; and seeps from 100-BC, 100-K, 100-N, 100-D, 100-H, 100-F), the Hanford Town Site (sediments from Hanford slough, Savage Island, and Locke Island; river water; and seep water from Hanford Spring); 300 Area (river water and springs), and downstream (sediments from McNary Dam and river water from the Richland Pumphouse).

**Table D.9. Maximum Detected Concentrations Evaluated for Aquatic Biota Dose Assessment**

Location Group	Radionuclide	Maximum Sediment (pCi/g)*	Maximum Water (pCi/L)*
Upstream	Hydrogen-3	—†	19
	Cesium-137	0.246	—
	Plutonium-239/-240	0.01	—
	Uranium-234	1.34	0.28
	Uranium-235	0.0996	0.0578
	Uranium-238	1.09	0.204
100 Areas	Hydrogen-3	—	5840
	Carbon-14	—	341
	Cobalt-60	0.0617	—
	Strontium-90	0.03‡	16.6
	Technetium-99	—	1.95
	Cesium-137	0.294	—
	Plutonium-239/-240	0.00376	—
	Uranium-234	1.09	0.297
Hanford Townsite	Uranium-235	0.115	0.0261
	Uranium-238	0.936	0.255
	Hydrogen-3	—	5750
	Cesium-137	0.269	—
	Plutonium-238	0.000549	—
	Plutonium-239/-240	0.00249	—
300 Area	Uranium-234	1.47	0.24
	Uranium-235	0.0923	0.0247
	Uranium-238	1.23	0.193
	Hydrogen-3	—	4350
Downstream	Uranium-234	—	30
	Uranium-235	—	2.75
	Uranium-238	—	29.7
	Hydrogen-3	—	24.9
	Technetium-99	—	0.621
	Cesium-137	0.259	—
Downstream	Plutonium-239/-240	0.00922	—
	Uranium-234	1.38	0.287
	Uranium-235	0.0954	0.033
	Uranium-238	1.32	0.24

\*1 pCi=0.037 Bq

†Not detected or not measured.

‡Although strontium-90 was not detected in sediment, the maximum non-detect was substantially less than the estimated concentration based on Kd.

Radionuclides were selected for the aquatic biota dose assessment based primarily on their detection in sediment or water and graphical comparisons of on-site to upstream and downstream sample results (Figures D.1 to D.29). In addition, having known or suspected sources from DOE operations, the results for tissue samples compared to reference (i.e. upstream, generally at Vernita Bridge), and the known potential for bioaccumulation was also used to identify which radionuclides should be included in the dose assessment. Most of the detected radionuclides in water – hydrogen-3 (tritium), carbon-14, strontium-90, technetium-99, and isotopic uranium – could readily be associated with known groundwater plumes. Most of the remainder of the detected radionuclides could have sources from DOE operations but due to

relatively high soil-water partition coefficients ( $K_d$ ) would most likely be associated with sediments instead of water. Cobalt-60, cesium-137, and isotopic plutonium were detected in sediments and may have sources from DOE operations. Although the magnitude of the sediment concentrations on site is sometime no greater than upstream, these radionuclides are included in the aquatic biota dose assessment. This is likely protective, as these radionuclides are not elevated above reference in tissues. Potassium-40 shows increased concentrations in sediment on site compared to both upstream and downstream levels. There were no differences in water or tissues potassium-40 concentrations between on-site and reference locations. There are also no groundwater plumes or other known DOE sources of potassium-40; therefore, potassium-40 was not included in the aquatic biota dose assessment.

As discussed in Section 4.2.6, biota doses were evaluated for Columbia River water and sediment and West Lake water and sediment, and soils (on site and off site). For West Lake, Tier 1 sum of fractions exceeded 1.0; therefore, Tiers 2 and 3 calculations were implemented using the mean water concentrations presented in Table D.10. The tiered screening process is further described in *A Graded Approach for Evaluating Radiation Doses to Aquatic and Terrestrial Biota* (DOE-STD-1153-2002).

**Table D.10. West Lake 2015 Water Samples**

Radionuclide	Water Concentration (pCi/L)		Sediment Concentration (pCi/g)	
	Maximum	Average	Maximum	Average
Cesium-137	—	—	0.335	0.317
Uranium-234	1650	393	4.14	4.12
Uranium-235	87.1	22.1	0.287	0.264
Uranium-238	1570	366	4.15	3.86
— Not detected or not measured. pCi=0.037 Bq.				

The Tier 1 and Tier 2 West Lake biota dose assessments were driven by the potential for dose from uranium isotopes in water and the assumed potential for these isotopes to accumulate in biota. Therefore, the Tier 3 West Lake biota dose calculations utilized site-specific information on bioaccumulation. As defined in DOE-STD-1153-2002, bioaccumulation is the ratio of the contaminant concentration in the organism relative to the contaminant concentration in an environmental medium resulting from the uptake of the contaminant from one or more routes of exposure. The more relevant biota data collected from West Lake are the brine flies sampled in 2000 and 2007 (PNNL-13487; DOE/RL-2007-50). Birds (avocets) were also sampled in 2000 and had lower concentrations than the brine flies (PNNL-13487). These birds are not year-round residents and thus have lower exposure and less potential for bioaccumulation at West Lake (DOE/RL-2007-50, Appendix K).

The maximum concentration of any of the uranium isotopes in brine flies was 0.88 pCi/g for uranium-233/-234 in 2007. The minimum uranium-233/234 West Lake pond water concentration was 940 pCi/L in 2007. The bioaccumulation factor is calculated by dividing the biota concentration (in pCi/g) by the water concentration (in pCi/mL), because RESRAD-BIOTA assumes that aquatic bioaccumulation occurs from water to biota. Therefore, the maximum bioaccumulation factor for uranium would be less than one (0.88 divided by 0.94). Also, as presented in Table D.10 of the 2011 report (DOE/RL-2011-119, Hanford Site Environmental Report for Calendar Year 2011), bioaccumulation factors for uranium isotopes based on the mean concentrations in flies and water were between 0.2 and 0.5. A bioaccumulation factor of one was used for the West Lake Tier 3 biota dose calculation as a somewhat protective measure of site-specific

uranium uptake into the food chain. This same approach was used in the 2011 (DOE/RL-2011-119) and 2012 (DOE/RL-2013-18, *Hanford Site Environmental Report for Calendar Year 2012*) reports. The data supporting the site-specific bioaccumulation factor are presented in those reports. Table D.11 presents the Tier 3 biota concentration guides for isotopic uranium for both aquatic and riparian animals. These site-specific values were used in the RESRAD-BIOTA Tier 3 screening discussed in Section 4.2.6.

**Table D.11. Tier 3 Biota Concentration Guides Calculated using RESRAD-BIOTA v1.5**

Radionuclide	Water BCG (pCi/L)*		Sediment BCG (pCi/g)*	
	Aquatic Animal	Riparian Animal	Aquatic Animal	Riparian Animal
Uranium-234	202000	20200	3030000	5270
Uranium-235	217000	21700	110000	3790
Uranium-238	222000	22200	42900	2490

\*1 pCi=0.037 Bq

Dose to terrestrial biota were evaluated using the near-field and far-field soil sample results. Table D.12 lists the maximum concentrations of strontium-90, cesium-137, plutonium-238, and plutonium-239/-240. These radionuclides were selected for the terrestrial biota dose assessment based on their detection in soil and graphical/statistical comparisons of soil/vegetation near-field to far-field sample results (Figures D.30 to D.40). In addition, having known or suspected sources from DOE operations, vegetation sample compared to soil results and the known potential for bioaccumulation were used to include or exclude radionuclides for the biota dose assessment.

**Table D.12. Maximum Detected Concentrations Evaluated for Terrestrial Biota Dose Assessment**

Location Group	Radionuclide	Maximum Soil (pCi/g)*
Near Field	Strontium-90	0.44
	Cesium-137	0.0016
	Plutonium-238	0.023
	Plutonium-239/240	0.053
Far Field	Strontium-90	14
	Cesium-137	0.14
	Plutonium-238	0.85
	Plutonium-239/240	1

\*1 pCi=0.037 Bq

A summary of the basis for selection of radionuclides is provided in the following:

- Strontium-90 and cesium-137 soil and vegetation concentrations in near-field samples were statistically greater compared to far-field samples.
- Isotopic plutonium soil concentrations in near-field samples were statistically greater than far-field samples.
- Near-field potassium-40 soil and vegetation concentrations were not different from far-field concentrations. There are also no known DOE sources of potassium-40; therefore, potassium-40 was not included in the terrestrial biota dose assessment.

- There were no statistical differences of antimony-125, cesium-134, europium-155, and isotopic uranium near-field concentrations compared to far-field concentrations; therefore, these radionuclides were not included in the terrestrial biota dose assessment.

The biota dose assessment also included supplemental calculations using measured concentrations in tissues. Presented in Section 4.2.6, these supplemental calculations provide a more realistic estimate of biota dose compared to doses calculated using the protective bioaccumulation assumptions made in the Tier 1 dose assessments. Section 4.2.6 lists the detected tissue concentrations evaluated in these supplemental dose calculations. The tissue concentrations are used for the internal dose calculations, whereas external dose is estimated from the relevant soil, sediment, and water concentrations. Where available, mean concentrations rather than maxima were used in the supplemental external dose calculations to be consistent with DOE guidance for these Tier 3 dose calculations. Note that radionuclides were detected in mule deer bone and mulberry fruit and leaf tissues. However, the detection in bone was only from a reference site, and the detections in mulberry were similar on-site and reference locations. Detections of radionuclides in mule deer and mulberry were therefore not included in the supplemental dose calculations.

Appendix D: Dose Calculations

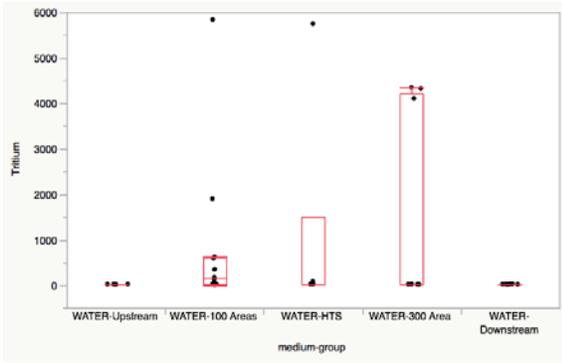


Figure D.1. Hydrogen-3 (tritium) concentrations in water

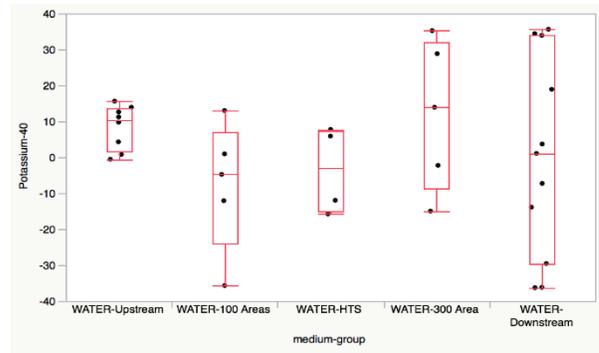


Figure D.5. Potassium-40 concentrations in water

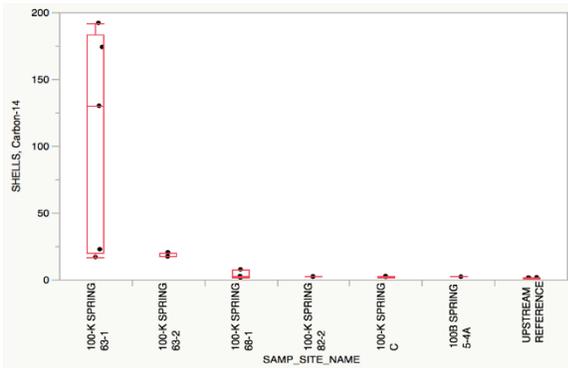


Figure D.2. Carbon-14 concentrations in clam shells

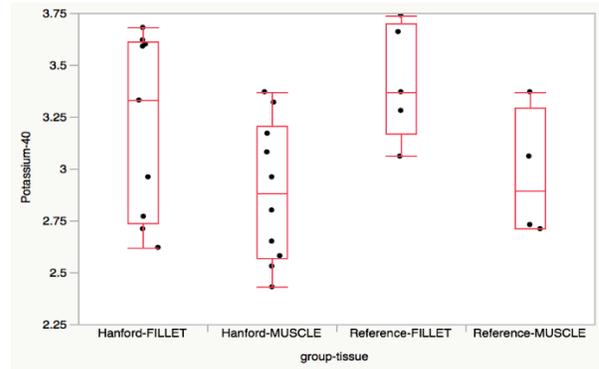


Figure D.6. Potassium-40 concentrations in tissues

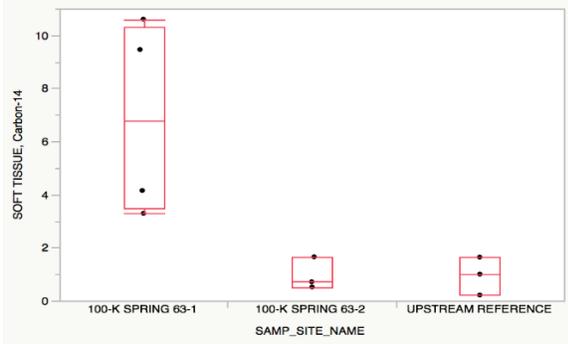


Figure D.3. Carbon-14 concentrations in clam shells

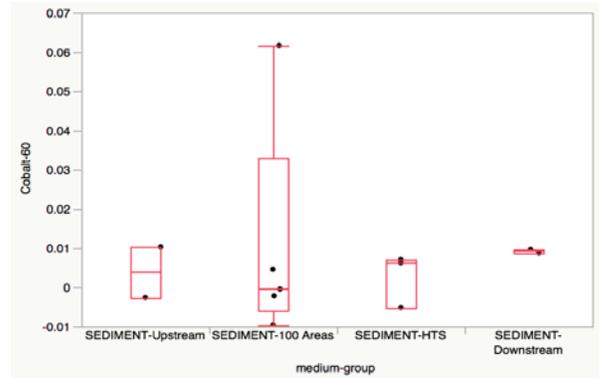


Figure D.7. Cobalt-60 concentrations in sediment

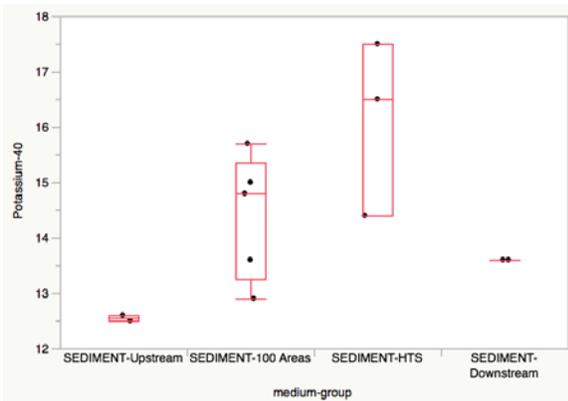


Figure D.4. Potassium-40 concentrations in sediment

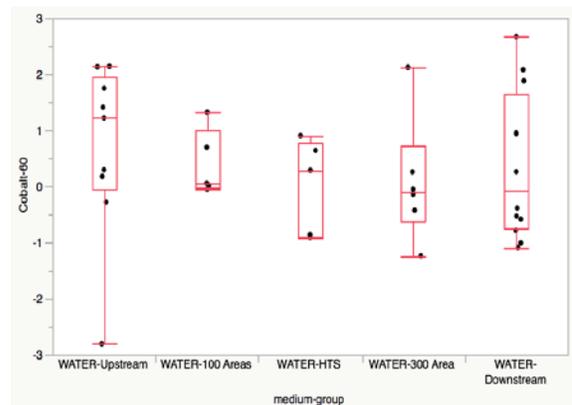


Figure D.8. Cobalt-60 concentrations in water

Appendix D: Dose Calculations

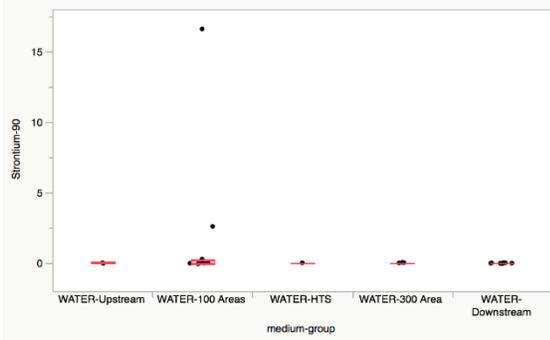


Figure D.9. Strontium-90 concentrations in water

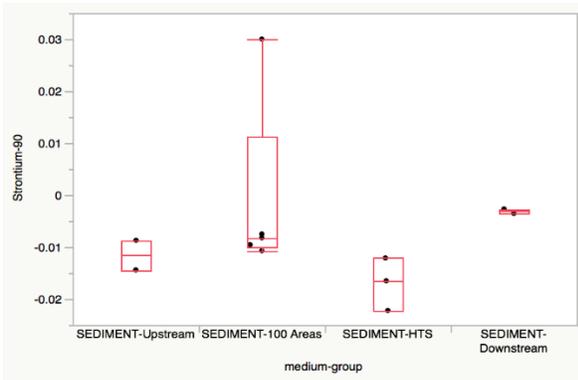


Figure D.10. Strontium-90 concentrations in sediment

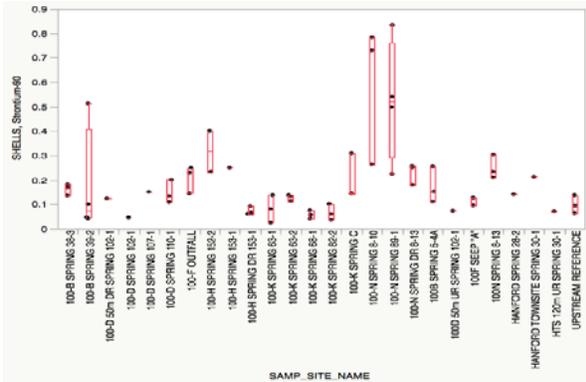


Figure D.11. Strontium-90 concentrations in clam shell

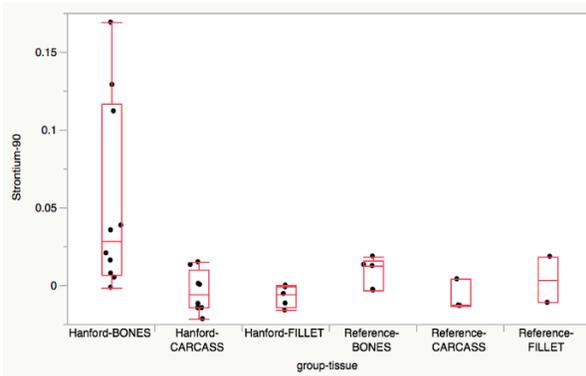


Figure D.12. Strontium-90 concentrations in tissues

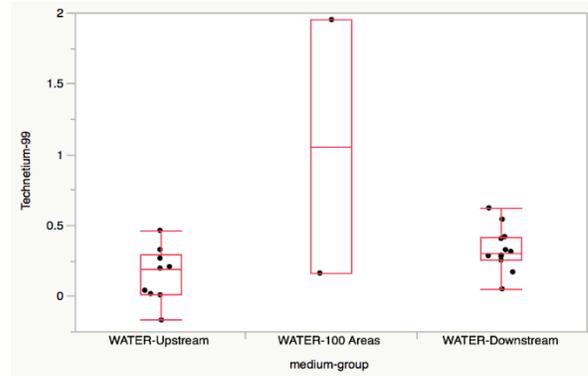


Figure D.13. Technetium-99 concentrations in water

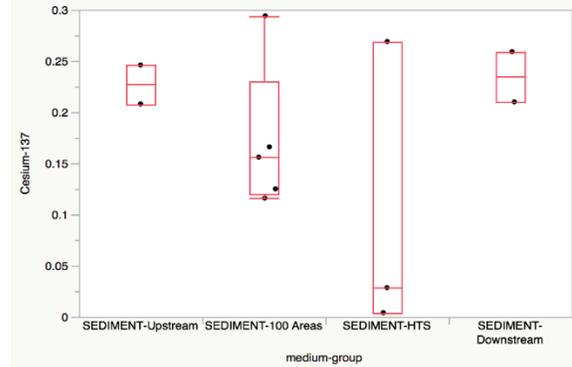


Figure D.14. Cesium-137 concentrations in sediment

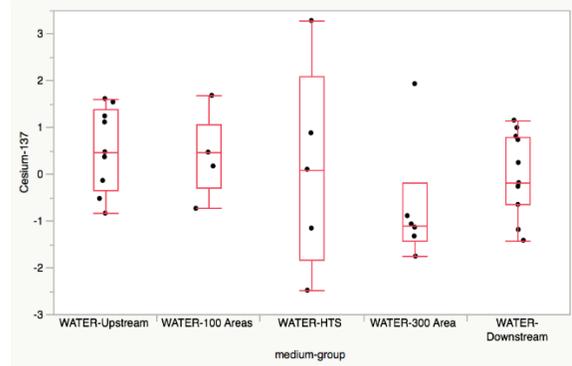


Figure D.15. Cesium-137 concentrations in water

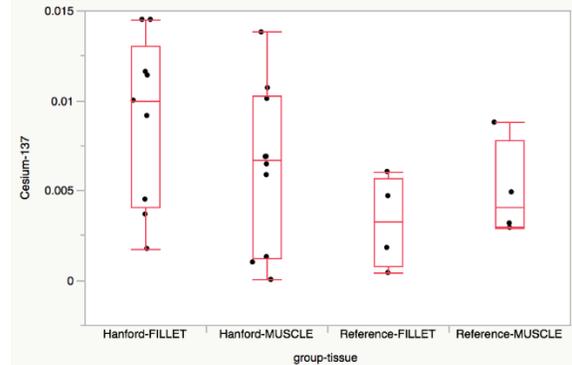


Figure D.16. Cesium-137 concentrations in tissues

Appendix D: Dose Calculations

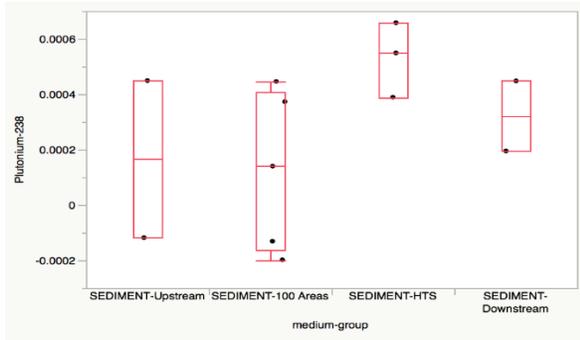


Figure D.17. Plutonium-238 concentrations in sediment

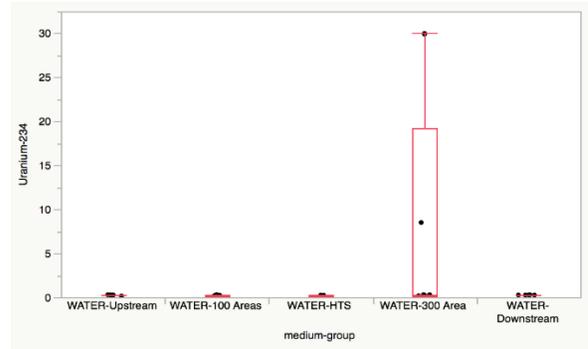


Figure D.21. Uranium-234 concentrations in water

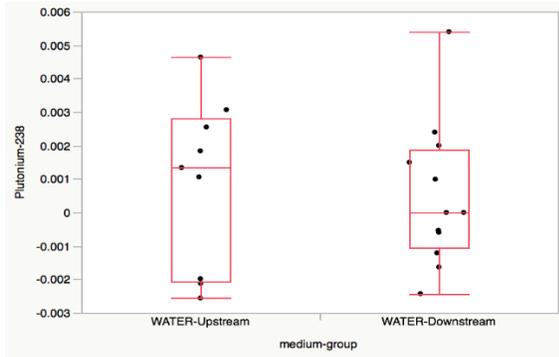


Figure D.18. Plutonium-238 concentrations in water

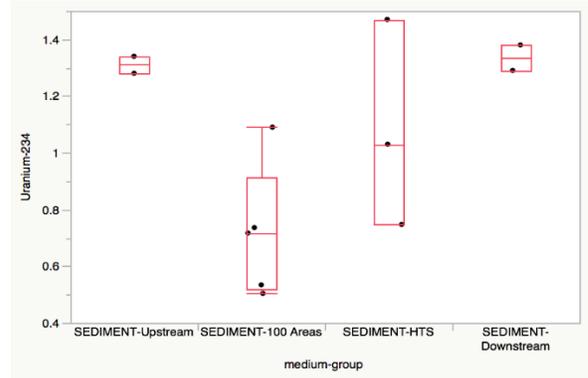


Figure D.22. Uranium-234 concentrations in sediment

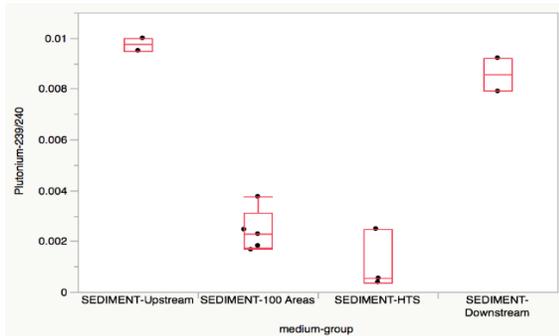


Figure D.19. Plutonium-239/240 concentrations in sediment

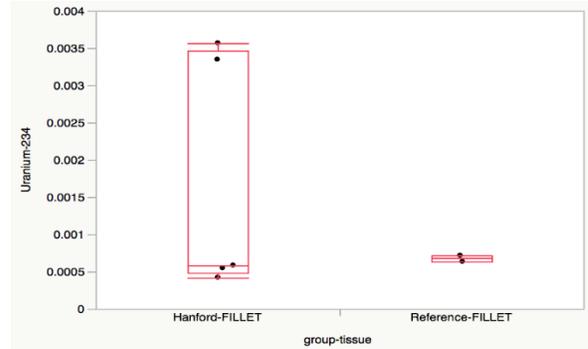


Figure D.23. Uranium-234 concentrations in tissues

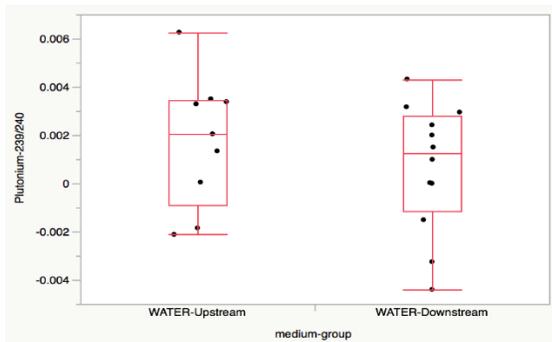


Figure D.20. Plutonium-239/240 concentrations in water

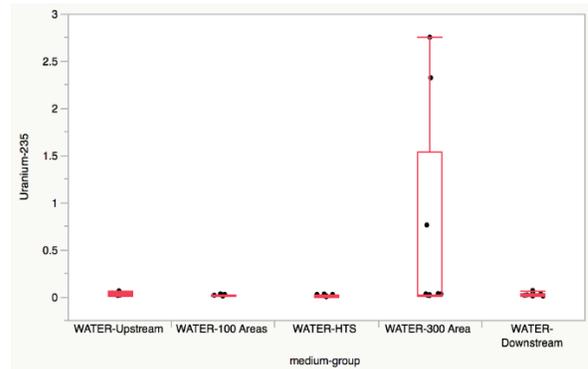


Figure D.24. Uranium-235 concentrations in water

Appendix D: Dose Calculations

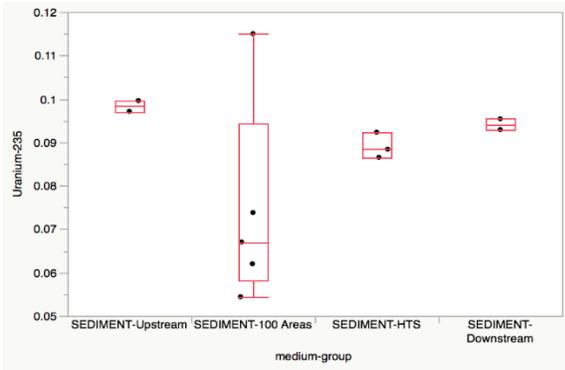


Figure D.25. Uranium-235 concentrations in sediment

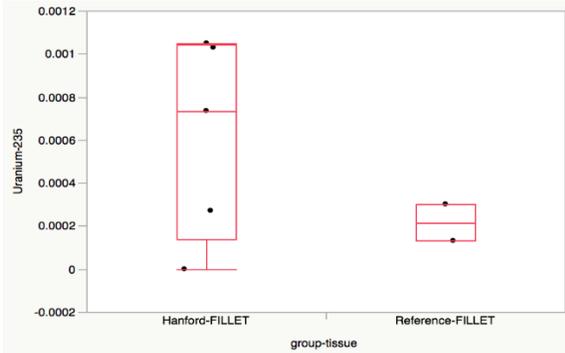


Figure D.26. Uranium-235 concentrations in tissues

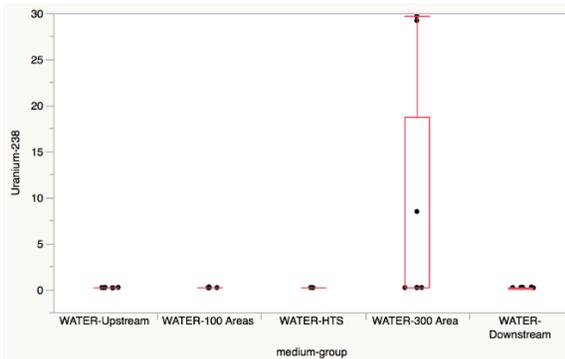


Figure D.27. Uranium-238 concentrations in water

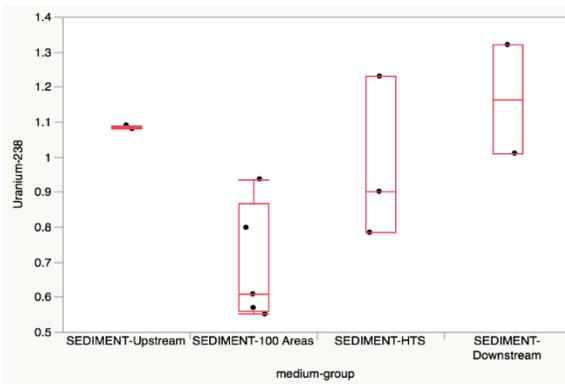


Figure D.28. Uranium-238 concentrations in sediment

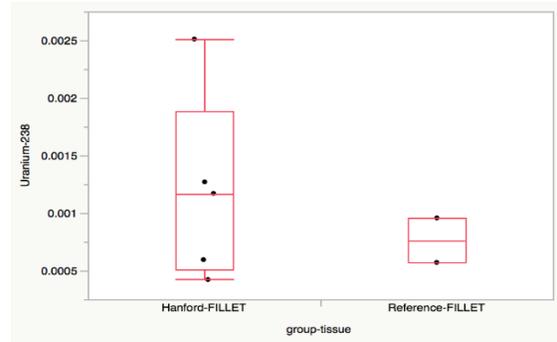


Figure D.29. Uranium-238 concentrations in tissues

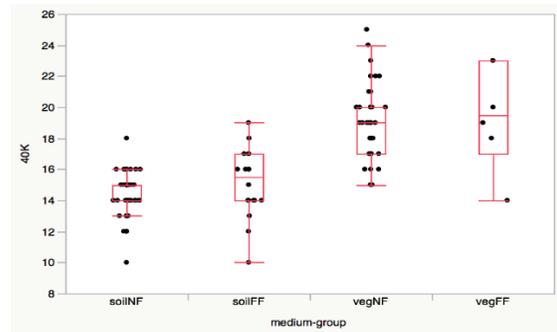


Figure D.30. Potassium-40 concentrations in soil and vegetation (veg) for near facility (NF) and far field (FF) samples

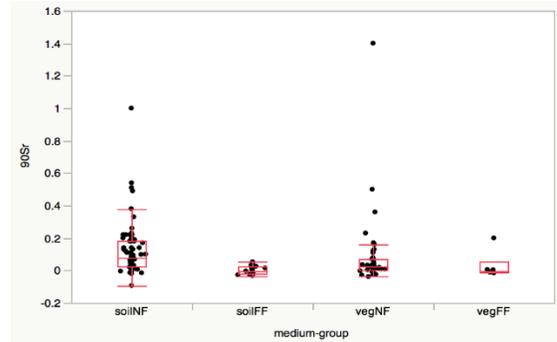


Figure D.31. Strontium-90 concentrations in soil and veg for NF and FF samples

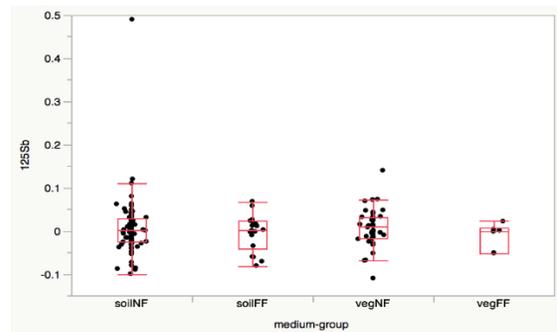


Figure D.32. Antimony-125 concentrations in soil and veg for NF and FF samples

Appendix D: Dose Calculations

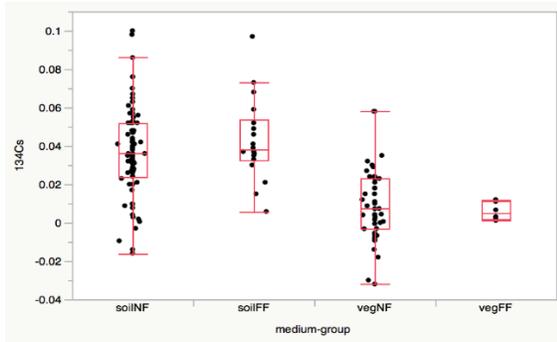


Figure D.33. Cesium-134 concentrations in soil and veg for NF and FF samples

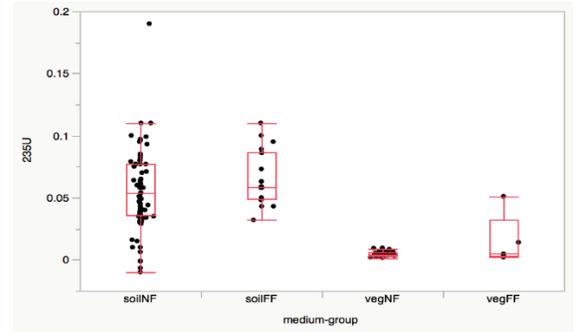


Figure D.37. Uranium-235 concentrations in soil and veg for NF and FF samples

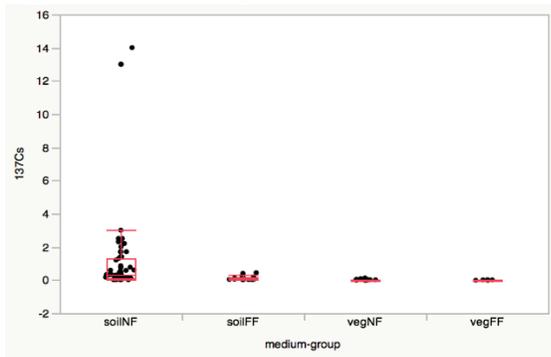


Figure D.34. Cesium-137 concentrations in soil and veg for NF and FF samples

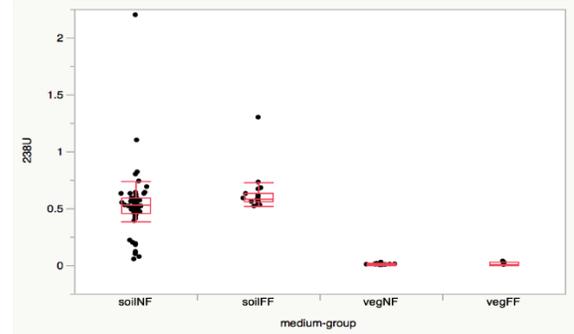


Figure D.38. Uranium-238 concentrations in soil and veg for NF and FF samples

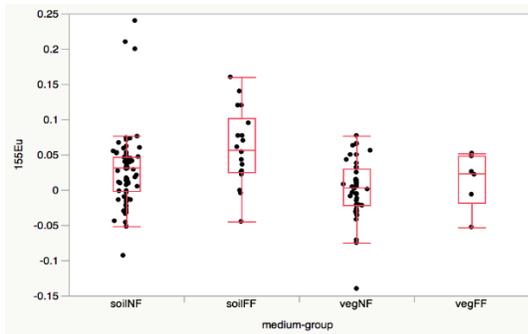


Figure D.35. Europium-155 concentrations in soil and veg for NF and FF samples

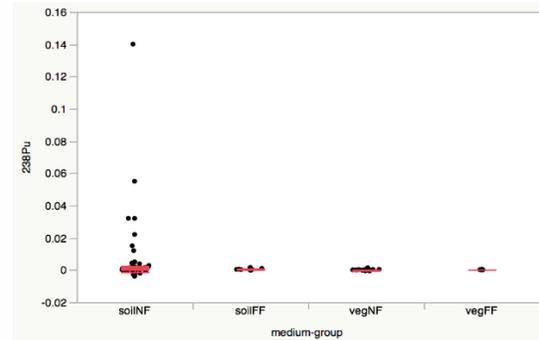


Figure D.39. Plutonium-238 concentrations in soil and veg for NF and FF samples

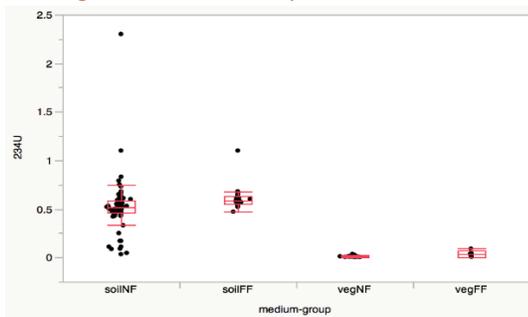


Figure D.36. Uranium-234 concentrations in soil and veg for NF and FF samples

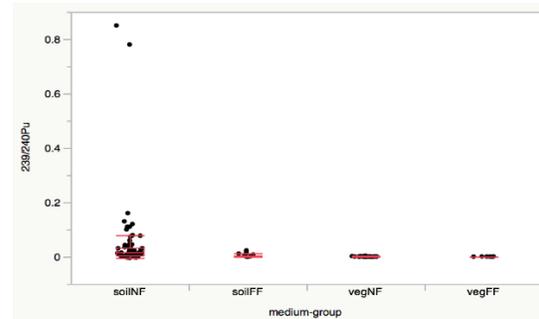


Figure D.40. Plutonium-238/240 concentrations in soil and veg for NF and FF sample